



Bioconversion of ecotoxic dehydroabietic acid using *Rhodococcus* actinobacteria

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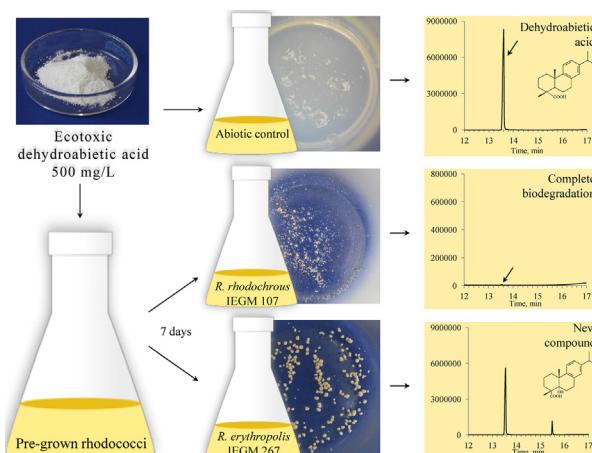
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HIGHLIGHTS

- Actinobacteria are capable to bioconversion of 500 mg/L dehydroabietic acid (DHA).
- The addition of DHA leads to increased rhodococcal respiration.
- DHA induces morphological-physiological changes in *Rhodococcus* actinobacteria.
- A novel metabolite of DHA bioconversion by *R. erythropolis* IEGM 267 was identified.

GRAPHICAL ABSTRACT



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ABSTRACT

Actinobacterial strains *Rhodococcus erythropolis* IEGM 267 and *R. rhodochrous* IEGM 107 were used to study biodegradation of dehydroabietic acid (DHA), a toxic tricyclic diterpenoid. The experiments were carried out in batch cultures of pre-grown rhodococci in the presence of 0.1% (v/v) *n*-hexadecane under aerobic conditions for 7 days. It was shown that *R. erythropolis* IEGM 267 and *R. rhodochrous* IEGM 107 partially and completely degraded DHA (500 mg/L), respectively. Characteristic physicochemical (reduced zeta potential) and morphological-physiological (increased average size of single cells and cell aggregates, increased root-mean-square roughness) changes in DHA-exposed actinobacteria were revealed. Products of DHA bioconversion by *R. erythropolis* IEGM 267 were analyzed and exhibited a previously unidentified metabolite 5 α -hydroxy-abeta-8,11,13-triene-18-oat. The obtained experimental data widen the knowledge on the catalytic activity of rhodococci and their possible contribution to decontamination of natural ecosystems from pollutants.

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1. Introduction

Resin acids are toxic tricyclic diterpenoids produced by coniferous trees of *Pinaceae* [1]. Those acids may account for up to 21.4% of the polar extractable fraction or 0.2–0.8% of the dry total weight of the wood [2]. A major component (up to 60%) of the resin fraction is dehydroabietic acid (DHA) 1 (Fig. 1) [3]. Mechanical and chemical pulping processes result in DHA release and in relatively high DHA concentrations (up to 500 mg/L) in the wastewater. Accumulated in natural ecosystems, DHA has a toxic effect on living organisms that can lead to ecological disbalances. It is established the acute toxicity towards various test organisms varies from 0.1 to 6.5 mg/L [1,4–6]. Consequently, to discover the effective methods to reduce the DHA concentration in the pulp and paper effluent is topical.

Due to natural bioaccumulation, conifer-produced DHA is found in marine and river water [7]; sediments and soil [8]; and living organisms [6]. In addition, the ability to produce DHA is customary for some representatives of streptomycetes and cyanobacteria [9,10]. The crucial role in natural detoxification of pollutants belongs to destructor microorganisms. Because DHA is widespread in the environment, microorganisms of various phylogenetic groups (fungi, bacteria) with the ability to partially or completely degrade DHA [1,11].

Biological treatment of wastewater based on enzymatic activity of microorganisms and natural processes of pollutant degradation is currently considered to be an advanced method which does not require aggressive reagents, and biological reactions proceed in one technological stage [1]. In most environments, DHA are found in low concentrations [11]. Therefore, many reported microorganisms generally exhibit degradation activity at DHA concentration not exceeding 250 mg/L [12–17] in the culture medium. Of the bacterial isolates characterized so far, most are gram-negative *Proteobacteria* with representatives in α -, β - and γ -subclasses. Their biotechnological application is potentially dangerous because of possible pathogenicity. The gram-positive biodestrutors are represented by few *Mycobacterium* and *Bacillus* strains [18–20]. According to the above, it is necessary to discover effective non-hazardous biocatalysts that retain effective degradation ability against higher DHA concentrations. One of the intensively studied groups of microorganisms in terms of biotechnological application are actinobacteria capable of transforming and degrading complex recalcitrant organic compounds at high concentrations [21,22]. Non-mycelial growth, lack of pathogenic properties, a flexible metabolic system and high oxygenase activity, all these determine the prospects for actinobacteria to be used as perspective biocatalysts [23–25]. At present, there are only a few studies describing actinobacteria in resin acid bioconversion [26,27]. This article reports on possible usage of actinobacteria from the Regional Specialised Collection of Microorganisms (IEGM; WDCM 768; <http://www.iegmcoll.ru/>) in effective DHA biodegradation.

2. Materials and methods

2.1. Microorganisms

In this work, 115 strains from the Regional Specialized Collection of Alkanotrophic Microorganisms representing the five species *Gordonia rubripertincta* (32 strains), *G. terrae* (21 strains), *R. erythropolis* (26 strains), *R. qingshengii* (3 strains), *R. rhodochrous* (3 strains) and *R. ruber* (30 strains) (Table 1) were used.

2.2. Chemicals

DHA (99.1%) was purchased from the Mosinter group limited (China). Ethyl acetate, hexane, and acetonitrile were of the

Table 1
Collection strains used in the work.

Species	No. of strains	Strain number in the IEGM collection
<i>G. rubripertincta</i>	32	95, 96, 97, 98, 100, 101, 103, 104, 105, 106, 109, 110, 118, 119, 120, 121, 122, 124, 126, 127, 128, 132, 133, 134, 135, 138, 139, 140, 723, 724, 725, 730
<i>G. terrae</i>	21	130, 136, 143 ^T , 144, 145, 146, 150, 151, 152, 153, 154, 155, 156, 157, 158, 159, 160, 161, 162, 163, 164
<i>R. erythropolis</i>	26	7 ^T , 11, 12, 14, 18, 19, 20, 21, 22, 23, 24, 25, 26, 98, 179, 180, 181, 182, 183, 184, 185, 186, 187, 188, 189, 267
<i>R. qingshengii</i>	3	1016 ^T , 247, 550
<i>R. rhodochrous</i>	3	64, 66, 107
<i>R. ruber</i>	30	70 ^T , 74, 80, 81, 82, 83, 84, 85, 86, 87, 88, 90, 91, 92, 93, 94, 172, 219, 220, 221, 222, 223, 224, 225, 226, 227, 228, 229, 230, 231

highest commercially available grade (Cryochrom, Russia). Pure *n*-hexadecane was purchased from Reachim (Russia). (Trimethylsilyl)diazomethane solution 2.0 M in diethyl ether was purchased from Sigma Aldrich (USA).

2.3. Use of DHA as the sole carbon source

Batch cultivations of actinobacteria were performed in 250 mL Erlenmeyer flasks containing 100 mL of the medium with shaking (160 rpm) at 28 °C. The mineral medium contained (per L) KNO₃ 1.0 g, K₂HPO₄ 1.0 g, KH₂PO₄ 1.0 g, NaCl 1.0 g, MgSO₄ × 7H₂O 0.2 g, CaCl₂ × 2H₂O 0.02 g, FeCl₃ 0.001 g, and 0.1% (v/v) trace element solution (1.5 g/L FeCl₃ × 7H₂O, 0.1 g/L H₃BO₃, 0.01 g/L ZnSO₄ × 7H₂O, 0.05 g/L Co(NO₃)₂ × 6H₂O, 0.005 g/L CuSO₄ × 5H₂O, and 0.005 g/L MnCl₂ × 4H₂O). DHA concentrations of 20, 40, 60, 80, and 100 mg/L as a powder or a solution in ethanol (1:10 mg/μL) were added as a carbon source. The medium was sterilized at 121 °C and 1.06 kg/cm² for 30 min. Suspensions of actinobacterial cells (OD₆₀₀ 1.0) pre-grown on agar (Oxoid, UK) for 2 days were used to inoculate media. The DHA biodegradation proceeded within 14 days. The sterile DHA-containing mineral medium was used as an abiotic control.

2.4. Screening test

In screening experiments on the ability of the collection strains to DHA biodegradation, an inoculum of a specific test strain and a mineral medium (according to paragraph 2.3) with 0.1% (v/v) yeast extract (10%) (Microgen, Russia) and 0.02–0.1% (v/v) *n*-hexadecane were used. DHA (500 mg/L) was dissolved in ethanol (1:10 mg/μL) and added after 48 h of cell cultivation. The DHA biodegradation proceeded for 7 days. The controls included (1) sterile DHA-containing mineral medium to evaluate the abiotic DHA stability; and (2) *n*-hexadecane- and bacterial cell-containing medium without DHA to differentiate possible bacterial metabolites.

2.5. Cell viability tests

To determine the number of living cells, the iodonitrotetrazolium violet (IV) staining method was used according to the following procedure: 100 μL of the test sample and 50 μL of 0.2% aqueous solution of IV were added in 96-well round-bottomed polystyrene microplates (Medpolymer, Russia). The IV was reduced to water insoluble formazan and this resulted in the development of red-violet staining, indicating actively respiring bacterial cells present in the sample. Samples were incubated at 28 °C for 24 h. After that period, the OD of the colored suspension was measured

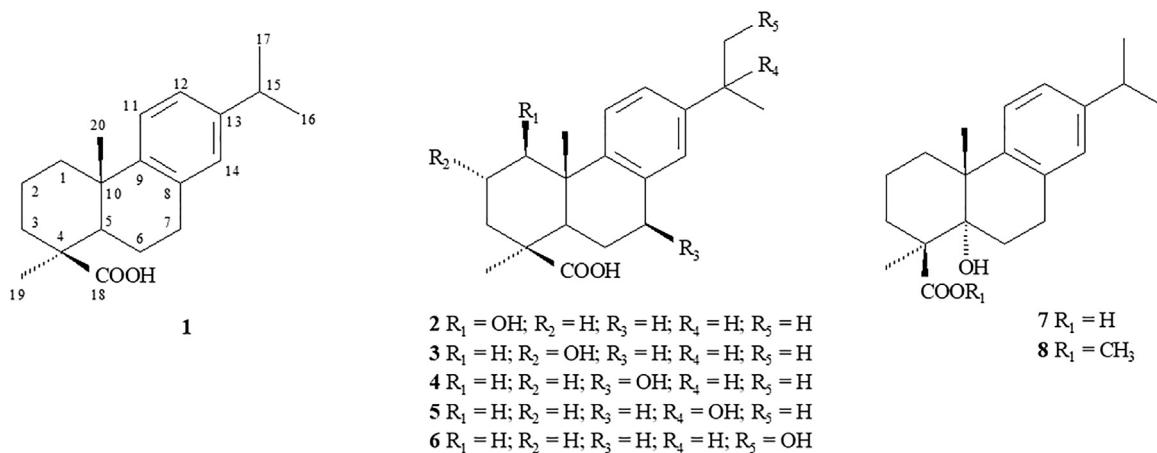


Fig. 1. Structures of DHA and DHA derivatives.

with a Multiscan Ascent microplate reader (Thermo Electron Corporation, Finland) at 630 nm wavelength [28].

2.6. Minimal inhibition concentration of DHA

Minimal inhibition concentration (MIC) of DHA against actinobacteria was determined by the micromethod of serial two-fold dilutions using 96-well round-bottomed polystyrene microplates. 100 μL of a sterile medium (LB) were inoculated into microplate wells. 100 μL of ethanol-dissolved DHA (working solution concentration, 10:100 mg/ μL) were added to the first well of each row and mixed thoroughly. 100 μL of the solution from the resulting mixture was transferred to the next well. The procedure was repeated until a series of two-fold dilutions was formed. The DHA concentration in one row ranged from 0.7 mg/L to 50 g/L. 10 μL (5×10^5 cells/mL) of the test culture pre-grown in LB were added to the resulting mixture. The plates were incubated at 28 °C for 3 days.

2.7. Phase contrast microscopy

Phase contrast microscopy was performed using an Axiostar plus light microscope (Carl Zeiss, Germany). Microscopic images and measurements of cell aggregate sizes were obtained with a PRO-150ES digital camera (Pixera, USA) and the VideoTest, Size 5.0 software (St. Petersburg, Russia).

2.8. Combined confocal laser and atomic force scanning microscopy

The morphology of bacterial cells was analyzed using a combined microscopic system which consists of the Olympus FV1000 laser confocal scanning microscope (CLSM) (Olympus Corporation, Japan) and the Asylum MFP-3D-BIO atomic force microscope (AFM) (Asylum Research, USA). Prior to imaging, a drop (15–20 μL) of the cell suspension was placed on a glass cover slip, mixed with the same volume of LIVE/DEAD® BacLight™ Bacterial Viability Kit (Invitrogen, USA) and allowed drying in darkness for 10–15 min. Then the cover slip was rinsed with deionized water and CLSM-scanned using an Olympus 100 \times oil immersion lens (numerical aperture, 1.4). SYTO 9 and propidium iodide included in LIVE/DEAD® were excited with the argon laser ($\lambda = 488$ nm) with a 505/525 nm barrier filter and the He-Ne laser ($\lambda = 543$ nm) with a 560/660 nm barrier filter, respectively. Image sizes of 0.12 × 0.12 mm (resolution, 1600 × 1600 pixels) were obtained at a speed of 40 nm/pixels. The images were analyzed by FV10-ASW 3.1 (Olympus Corporation, Japan). The cell volume and area were

additionally calculated according to Neumann et al. [29]. The cell volume was calculated by the formula:

$$V = r^2\pi h(\mu m^3),$$

cell surface area:

$$a = 2r^2\pi + \pi rh(\mu m^2),$$

where r – ½ cell width, h – cell length, $\pi = 3.14$.

The CLSM image was imported into the Igor Pro 6.22A (WaveMetrics, USA) software used for AFM. AFM scanning of the preparations was carried out in a semi-contact mode in air. Silicon cantilevers AC240TS (Olympus, Japan) with resonance frequencies of 50–90 kHz and spring constants of 0.5–4.4 N/m were used.

2.9. Zeta potential measurement

Electrokinetic (zeta) potential of bacterial cells was measured by the dynamic light scattering using a ZetaSizer Nano ZS analyzer (Malvern Instruments, UK) with Malvern ZetaSizer, v. 2.2. The actinobacterial suspension grown with 0.1% *n*-hexadecane and DHA (500 mg/L) was washed twice and resuspended in 10 mM KNO₃ (pH 7.0) to achieve OD₆₀₀ 0.2.

2.10. Respirometry analysis

Respiratory activities of bacterial cells were measured using a six-channel indirect cycle MicroOxymax® respirometer (Columbus Instruments, USA) connected to a PC. Experiments were performed in 300 mL Micro-Oxymax bottles. The bacterial cultures were constantly stirred (300 rpm, 28 ± 2 °C) on a multi-position magnetic stirrer RT 10 (Power IKAMAG, Germany). The amount and rate of O₂ consumed and CO₂ released (μL) were evaluated. The parameters of respiratory activity were recorded automatically every 42 min for 9 days.

2.11. DHA derivative isolation and identification

To isolate residual DHA and its metabolites, the culture media were acidified with 10% HCl solution and extracted three times with an equal volume of ethyl acetate. The combined organic layers were washed with an aqueous solution of 1% Na₂CO₃ and then with distilled water to a final pH of 7.0. The ethyl acetate extract was dried over anhydrous Na₂SO₄. The solvent was removed with a rotary evaporator (Heidolph, Germany). A mixture of DHA metabolites was methylated with (trimethylsilyl)diazomethane for 1 h. DHA biodegradation products were qualitatively analyzed by TLC on sil-

ica gel 60 plates (Merk, Germany). The sample spots were visualized by treating plates with ethyl acetate:*n*-hexane (3:7, v/v) and spraying with a solution of 5% H₂SO₄ followed by heating at 95–100 °C for 2–3 min. The GC–MS of the DHA biodegradation products were performed using an Agilent 6890N/5975B chromatograph (Agilent Technologies, USA) equipped with a HP-5MS capillary column (30 m × 0.25 mm, 0.25 µm) and operating in the electron impact ionization mode (70 eV). Helium was used as a carrier gas (1 mL/min). The column temperature was programmed from 120 to 320 °C. MS were recorded in the range of 40–460 *m/z* and compared with those from the NIST08 Library. The DHA biodegradation dynamic was quantitatively analyzed by HPLC using a LC Prominence 20 AD chromatograph (Shimadzu, Japan) fitted with a reversed-phase column Supelcosil™ LC-18 (150 × 4.5 mm, 5 µm) and a diode matrix detector (SPD-M20A). The mobile phase contained 70% acetonitrile solution; the flow rate was 1 mL/min; the detection wavelength at 190 nm; column temperature was 40 °C. The retention time for DHA and compound **8** was 14.99 ± 0.02 and 7.88 ± 0.02 min, respectively. The data were registered and processed using the LCSolution program (v. 1.25).

2.12. Preparative isolation of methyl 5α-hydroxy-abiet-8,11,13-triene-18-oate (**8**) from the mixture of DHA bioconversion products by *R. erythropolis* IEGM 267

To methylate the DHA biodegradation products, the concentrated extract (3 g) was dissolved in 70 mL acetone, than methyl iodide (4.26 g) and K₂CO₃ (4.14 g) were added to the solution. The reaction mixture was boiled with a reflux condenser under stirring for 1.5 h. After the reaction was completed, the mixture was filtered and the solvent was removed with a rotary evaporator. The residue was separated using a flash-chromatograph (BÜCHI, Switzerland) on a cartridge 40 mm in diameter and 150 mm long (BÜCHI, Switzerland) containing 90 g of silica gel 60. The ratio of the methylated extract and the sorbent was 1:30. The elution was carried out with the hexane:ethyl acetate mixture using the concentration gradient from 99:1 to 9:1. The isolated yield of compound **8** after column chromatography was 17%.

The IR spectrum of the compound **8** dissolved in CHCl₃ was recorded on the Bruker 66/S IFS Fourier spectrometer (Bruker, Germany). The 1D and 2D NMR spectra of compound **8** dissolved in CDCl₃ were recorded on the Bruker AVANCE II (Bruker BioSpin GmbH, Germany, 400 and 100 MHz for ¹H- and ¹³C-NMR spectra, respectively). Chemical shifts (δ) were stated in units of parts per million (ppm) relative to the TMS as an internal standard. The optical rotation was measured on the Perkin Elmer 341 polarimeter (Perkin Elmer, USA) using sodium D light for CHCl₃ solution. The melting point was measured using the OptiMelt MPA100 (Stanford Research Systems, USA) instrument at the heating rate of 1 °C/min.

3. Results and discussion

3.1. Screening test

Many aerobic bacterial isolates were isolated from pulp and paper industry effluent, biological wastewater treatment systems and from bottom and soil samples taken in the immediate area of pulp and paper mills. These isolates use resin acids as the sole carbon and energy source [14,30–32]. It has been shown that only *R. erythropolis* IEGM 267 from all strains studied is capable to use DHA at a concentration of 20 mg/L as the sole carbon source. The remaining cultures did not exhibit any growth in the presence of 20–100 mg/L DHA within 14 days of the experiment. The ability of actinobacteria to utilize DHA in the presence of 0.5% (v/v) *n*-hexadecane as the inducer of ecotoxicant decomposition has been

previously described [26]. *n*-Hexadecane present in the culture medium promotes the induction of multi-purpose oxygenase systems in alkanotrophic actinobacteria [33]. Furthermore, due to the lack of the catabolic repression in actinobacteria, particularly in rhodococci, *n*-alkanes can be used as co-substrates in biotransformation processes of recalcitrant compounds [34]. In a number of studies on actinobacterial biotransformations of complex organic compounds, the *n*-hexadecane concentration of 0.1% (v/v) was proved to be optimal [35,36]. Screening of the actinobacterial catalytic activity against DHA was carried out under similar conditions, and DHA was added after 48 h of bacterial cell cultivation.

Ten strains capable of complete or partial degradation of DHA (Table 2) were selected. *R. rhodochrous* IEGM 107 possesses the greatest potential for natural system decontamination from resin acids. It catalyzed the high (>98%) DHA biodegradation. Five strains could degrade more than 50% of DHA. Their catalytic activities were distributed in the order of *G. rubripertincta* IEGM 104 > *G. terrae* IEGM 150 > *G. rubripertincta* IEGM 109 > *G. rubripertincta* IEGM 105 > *R. ruber* IEGM 80. TLC showed that *G. rubripertincta* IEGM 100, IEGM 120, IEGM 132 and *R. erythropolis* IEGM 267 catalyzed the DHA bioconversion to form a more polar product. According to GC–MS, the detected unknown metabolite gives a mass spectrum with a molecular-ion peak at *m/z* 330.3.

3.2. Study of the DHA biodegradation process using *R. rhodochrous* IEGM 107 and *R. erythropolis* IEGM 267

A detailed study of the DHA bioconversion process using the most active biocatalysts *R. rhodochrous* IEGM 107 (GenBank: MG585109) and *R. erythropolis* IEGM 267 (GenBank: MRBQ00000000.1) was carried out. It included sampling every 24 h for microscopic analyses, HPLC of residual DHA and its metabolites, viable cell determinations, measurements of respiratory activity and pH of the reaction medium.

As shown in Fig. 2A, a decrease in the viable cell number of *R. rhodochrous* IEGM 107 was observed during 48 h after DHA addition to the culture medium. Presumably, this experimental stage corresponded to the adaptation period of *R. rhodochrous* IEGM 107 to a DHA toxic effect, with its concentration decreased to 356 mg/L within 48 h. The next day marked a rapid increase in growth and catalytic activity of bacterial cells, with only 104 mg/L DHA observed in the culture medium. Then the DHA concentration decreased gradually, and the residual DHA was less than 5 mg/L on the day 9. Different dynamics of the viable cell number was observed during DHA biodegradation by *R. erythropolis* IEGM 267. As shown in Fig. 2B, the addition of DHA led to 1.5–2.8 times higher rhodococcus cell inhibition compared to the biotic control throughout the experiment. As revealed by HPLC, the gradual decrease in DHA concentration in the culture medium was accompanied by the metabolite *m/z* 330.3 formation, with its maximum (14%) concentration on day 8 of the experiment. The DHA effects on the viable cell number are consistent with MICs of DHA. For *R. rhodochrous* IEGM 107, the MIC value was 780 mg/L, indicating relatively high resistance to DHA toxic effects compared with many known biodestructors [30,31,37]. Conversely, *R. erythropolis* IEGM 267 were much less resistant to DHA toxic effect (MIC 24 mg/L).

According to phase-contrast microscopy analyses, rhodococci formed compact cellular aggregates of 10–50 µm in size with more than ten viable cells in the mineral medium in the presence of *n*-hexadecane. Subsequent addition of DHA resulted in the formation of cellular aggregates up to 0.5 mm in size for *R. rhodochrous* IEGM 107, while the size of *R. erythropolis* IEGM 267 aggregates achieved 5 mm.

The physiological state of actinobacteria in the presence of DHA was revealed by CLSM data. After 48 h of DHA addition, cells of *R. rhodochrous* IEGM 107 were still viable, but cells with red fluores-

Table 2DHA biodegradation by actinobacteria in the presence of *n*-hexadecane.

Strain	Products of biodegradation*, %		Concentration residual DHA**, mg/L
	DHA	Metabolite <i>m/z</i> 330.3	
Abiotic control	100.0	—	490.0 ± 4.0
<i>G. rubripertincta</i> IEGM 100	90.5 ± 5.2	9.5 ± 0.9	443.0 ± 25.4
<i>G. rubripertincta</i> IEGM 104	9.0 ± 0.5	—	44.1 ± 2.4
<i>G. rubripertincta</i> IEGM 105	44.0 ± 5.2	—	215.0 ± 25.4
<i>G. rubripertincta</i> IEGM 109	41.0 ± 4.8	—	200.9 ± 23.5
<i>G. rubripertincta</i> IEGM 120	98.8 ± 0.9	1.2 ± 0.5	484.1 ± 4.4
<i>G. rubripertincta</i> IEGM 132	92.2 ± 3.5	8.2 ± 3.5	451.7 ± 17.5
<i>G. terrae</i> IEGM 150	35.0 ± 3.7	—	171.5 ± 18.1
<i>R. erythropolis</i> IEGM 267	86.0 ± 4.9	14.0 ± 1.2	421.4 ± 24.0
<i>R. rhodochrous</i> IEGM 107	≤ 2.0	—	4.0 ± 2.0
<i>R. ruber</i> IEGM 80	49.0 ± 5.7	—	240.0 ± 27.9

The GC-MS* and HPLC** data after 7-day cultivation of actinobacteria in the presence of DHA (500 mg/L) are shown.

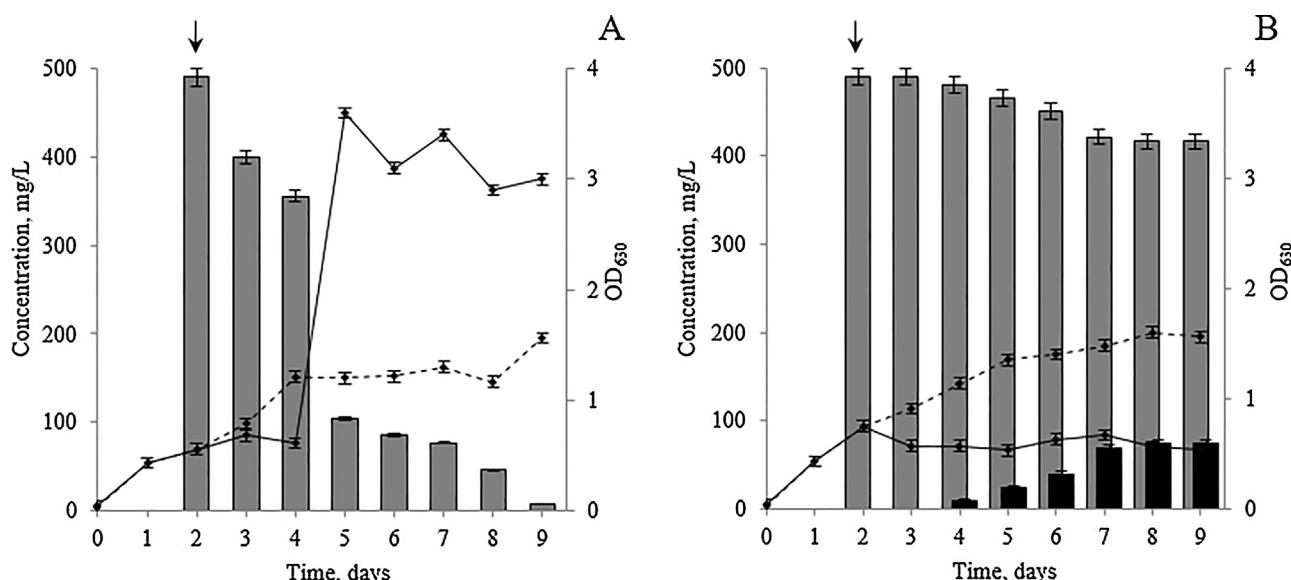


Fig. 2. Dynamics of DHA biodegradation by *R. rhodochrous* IEGM 107 (A) and *R. erythropolis* IEGM 267 (B). Concentrations of (■) DHA and (■) metabolite *m/z* 330.3 (mg/L), (—◆—) bacterial growth (*OD*₆₃₀) during DHA biodegradation, (—◆—) bacterial growth in the presence of *n*-hexadecane. The arrow indicates the time points when DHA was added.

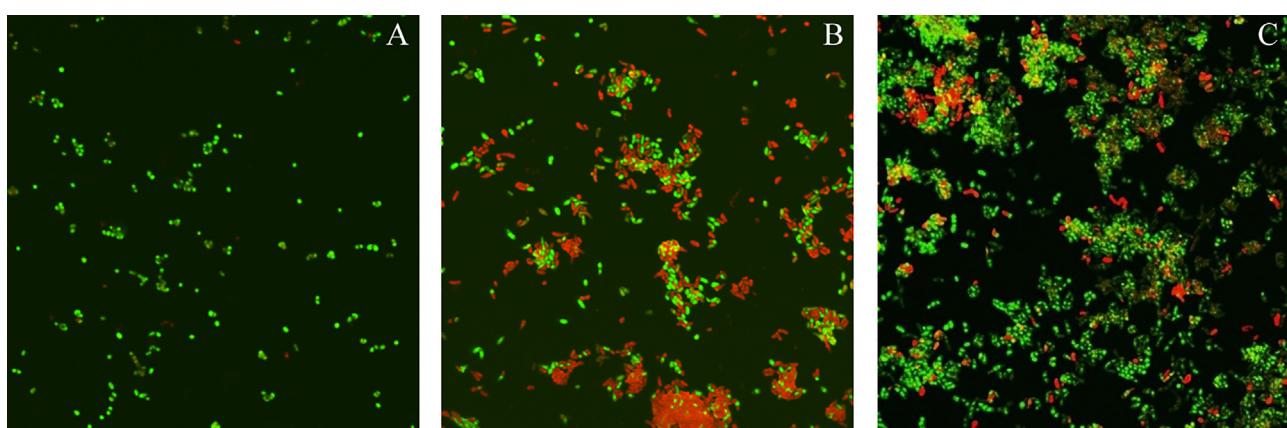


Fig. 3. CLSM images (100×) of *R. rhodochrous* IEGM 107 before DHA addition (A); after 48 h (B) and 72 h (C) of cultivation with DHA. Green cells are alive, red ones are dead. (For interpretation of the references to color in this figure legend, the reader is referred to the web version of this article.)

cence were present in the CLSM images, indicating probably some toxic effect of the resin acid. After 72 h of cultivation in the presence of DHA, the number of viable cells increased (Fig. 3). Probably, this effect was due to cell adaptation to the ecotoxicant.

After 24 h cultivation of *R. erythropolis* IEGM 267 in the presence of 500 mg/L DHA, the damaged cells along with live and dead cells (Fig. 4) were found in the medium. A characteristic "double" (green-orange) luminescence of certain parts of a cell or of the whole cell may indicate the damaged membrane of viable cells. Similar effect

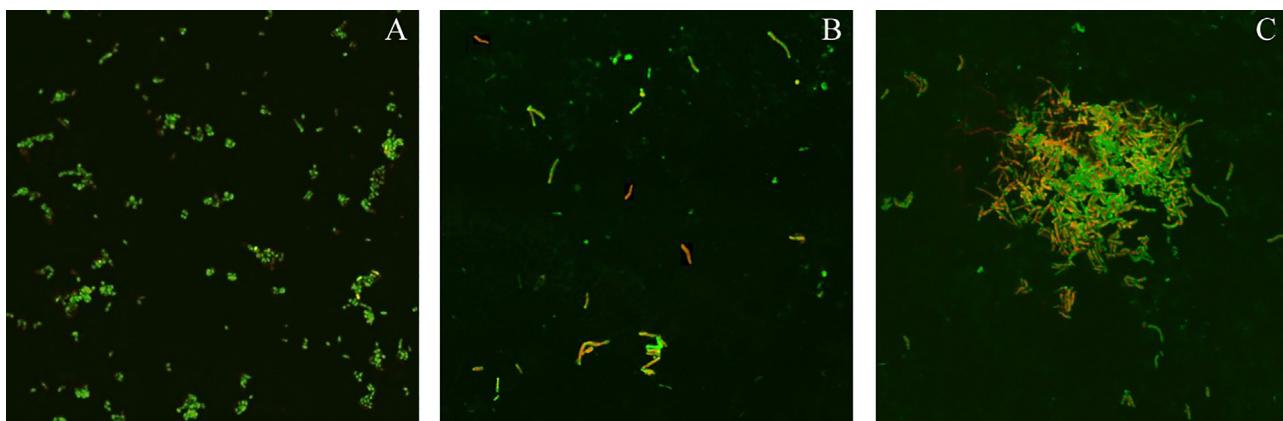


Fig. 4. CLSM images ($100\times$) of *R. erythropolis* IEGM 267 before DHA addition (A); after 24 h (B) and 168 h (C) of cultivation with DHA. Green cells are alive, red ones are dead. (For interpretation of the references to color in this figure legend, the reader is referred to the web version of this article.)

Table 3
DHA effects on morphometric parameters.

Conditions	Length, μm	Width, μm	Volume, μm^3	Area, μm^2
<i>R. rhodochrous</i> IEGM 107				
Mineral medium with <i>n</i> -hexadecane	1.38 ± 0.28	1.13 ± 0.19	1.38 ± 0.16	4.45 ± 0.28
Mineral medium with <i>n</i> -hexadecane plus DHA	1.82 ± 0.26	1.26 ± 0.24	2.26 ± 0.13	6.09 ± 0.23
<i>R. erythropolis</i> IEGM 267				
Mineral medium with <i>n</i> -hexadecane	3.10 ± 1.03	1.01 ± 0.09	2.65 ± 0.28	6.85 ± 0.18
Mineral medium with <i>n</i> -hexadecane plus DHA	6.83 ± 1.96	0.85 ± 0.11	3.87 ± 0.36	10.24 ± 0.24

Data are presented for living cells of actinobacteria after 7-day cultivation with 500 mg/L DHA and 0.1% (v/v) *n*-hexadecane.

(double luminescence) was observed when rhodococcal cells were stained with LIVE/DEAD dye after their exposure to high concentrations of organic solvents [38].

Morphometric studies using AFM analysis made it possible to record the average values of the cell size parameters. As shown in Table 3, the DHA additions induced changes in the cell size, in particular their increased length and width. The most pronounced size changes were observed when DHA was added to the culture medium of *R. erythropolis* IEGM 267. At the same time, an increase in the ratio of the cell surface area to its volume was noted, which, according to the literature data [29], promotes a more efficient cell-substrate contact.

AFM images of DHA-incubated *R. rhodochrous* IEGM 107 revealed extracellular fluid formation for many samples scanned (Fig. 5A). This phenomenon could not be a consequence of the lysis of bacterial cells because damaged cells were not found in the *Rhodococcus* population. When CLSM and AFM images were aligned, living cells were observed in the liquid. We assume that the extracellular fluid may be a biosurfactant or a mixture of a biosurfactant, *n*-hexadecane, and DHA. It is known, that biosurfactant synthesis contributes to the increased contact area of actinobacteria with hydrophobic compounds and their bioavailability [39]. Similar results were obtained by scanning actinobacterial cells during betulin biotransformation [40]. Atrat et al. suggested earlier that in β -sitosterol biotransformation by *Mycobacterium fortuitum* NRRL B-8119, a multicomponent mobile lipophilic mesophase was formed, which gradually dissolved and triggered the transport mechanism of the substrate into the cell [41]. Likewise, the AFM images of the obtained actinobacterial biosurfactant adsorbed on the glass exhibited micelles and vesicles formed around the bacterial cells [42]. Similar structures were seen in our images. AFM images of *R. erythropolis* IEGM 267 cells incubated with DHA showed no formation of extracellular fluid (Fig. 5B), while DHA granules were observed on the cell surface.

Table 4
Changes in the actinobacterial cell surface.

Culture conditions	Roughness, nm	Electrokinetic potential
<i>R. rhodochrous</i> IEGM 107		
Mineral medium with <i>n</i> -hexadecane	206.5 ± 10.7	-26.6 ± 0.9
Mineral medium with <i>n</i> -hexadecane plus DHA	365.9 ± 6.9	-27.3 ± 1.1
<i>R. erythropolis</i> IEGM 267		
Mineral medium with <i>n</i> -hexadecane	100.7 ± 9.8	-15.5 ± 1.4
Mineral medium with <i>n</i> -hexadecane plus DHA	289.3 ± 12.4	-19.8 ± 1.0

Data of AFM scanning and Zeta measurements of living actinobacterial cells after 7-day cultivation with 500 mg/L DHA plus 0.1% (v/v) *n*-hexadecane are shown.

DHA effects on actinobacteria were accompanied by changes in their cell surfaces. According to AFM, a characteristic feature of actinobacterial cell surface in the presence of DHA was increased root-mean-square roughness (Table 4). The electrokinetic (Zeta) potential data for bacterial cells suggest that addition of DHA has a more pronounced change in this parameter for *R. erythropolis* IEGM 267 (by 21.7%). A similar effect may be associated with a higher resistance of *R. rhodochrous* IEGM 107 to toxic effects of DHA compared to *R. erythropolis* IEGM 267 (MIC data).

Respiratory activity as one of the indicators of bacterial cell viability and intensity of metabolic processes was used to monitor DHA biodegradation in this study. As seen from Fig. 6A, DHA added to the culture medium on day 1 decreased the respiration rate. For *R. rhodochrous* IEGM 107, the decreased respiration rate correlated with the decreased number of viable cells. On day 5, the increased growth and also respiration rate was observed.

For *R. erythropolis* IEGM 267, the respiratory activity measurements during the DHA biodegradation showed that O_2 uptake and CO_2 release rates were independent of the number of viable cells over the experiment and correlated with the

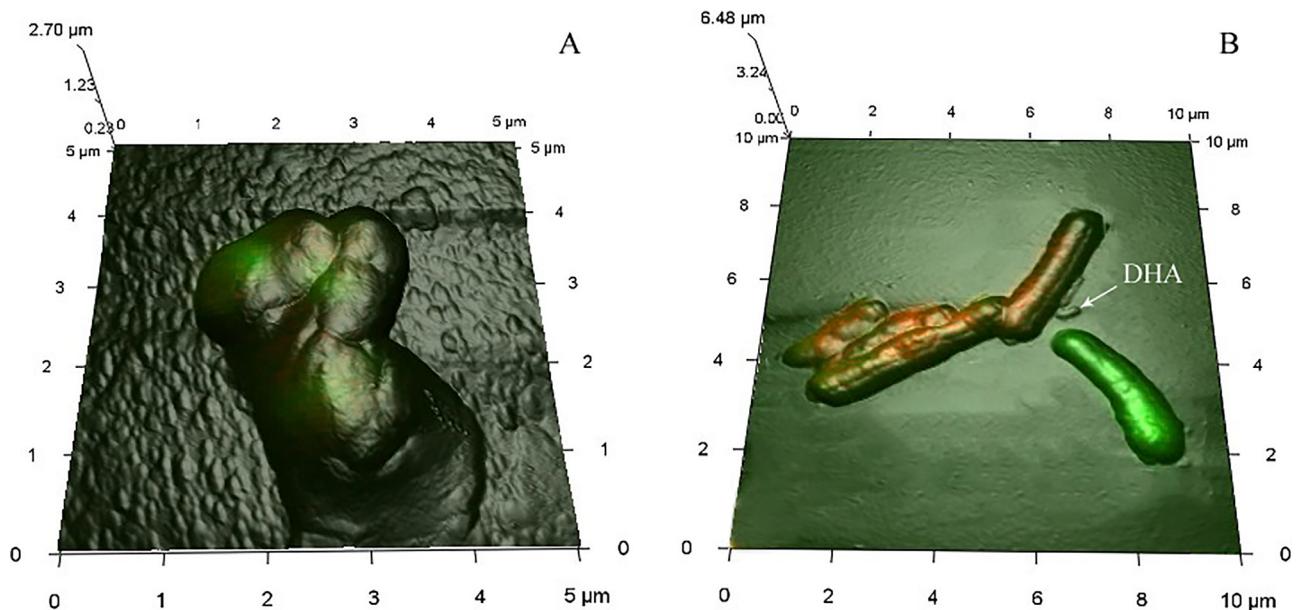


Fig. 5. AFM-CLSM images of *R. rhodochrous* IEGM 107 (A) and *R. erythropolis* IEGM 267 (B) after 168 h addition of DHA. The arrow indicates DHA granules.

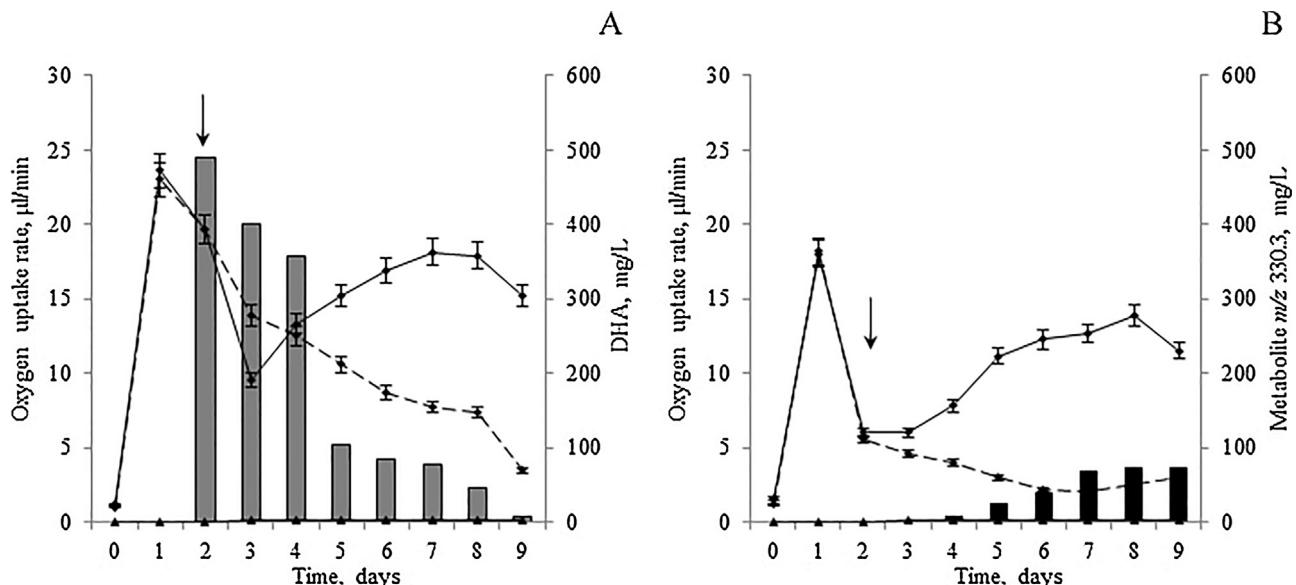


Fig. 6. Respiratory activities of *R. rhodochrous* IEGM 107 (A) and *R. erythropolis* IEGM 267 (B). Oxygen uptake during DHA biodegradation (—◆—), growth in the presence of *n*-hexadecane (—◆—), abiotic control (—▲—), concentrations of (■) DHA and (■) metabolite *m/z* 330.3 (mg/L). The arrow indicates the time when DHA was added.

formation-accumulation of the DHA metabolite. Indeed, the highest respiration rates (13.41–14.28 μL/min) were measured after DHA addition to the nutrient medium on day 8 of the experiment (Fig. 6B) when the maximum level of metabolite formation was observed (see Fig. 1B). In general, the total amount of CO₂ released and O₂ consumed within 9 days of DHA biodegradation by *R. erythropolis* IEGM 267 (139088 μL of CO₂ and 149941 μL of O₂, respectively) was 1.5 times lower than those with *R. rhodochrous* IEGM 107 (223959 μL of CO₂ and 216549 μL of O₂, respectively). At the same time, the average respiration rate for *R. rhodochrous* IEGM 107 in the experiment (15 μL/min) was higher than that of *R. erythropolis* IEGM 267 (10 μL/min).

As previously shown, the level of actinobacterial catalytic activity is influenced by the pre-growth of bacterial cells in a medium with *n*-hexadecane [26,35,36]. In this respect, we studied the

dependence of the metabolite *m/z* 330.3 formation on the initial concentration of *n*-alkane in the culture medium of *R. erythropolis* IEGM 267 (Fig. 7). According to our data, the addition of more than 0.1% (v/v) *n*-hexadecane led to a decreased degradation activity of rhodococci, with metabolite *m/z* 330.3 being not detected in the medium. At the same time, a noteworthy correlation of the growth and degradation activities with a decrease in the alkane concentration in the medium was observed. As shown in Fig. 7, rhodococci pre-grown with 0.06% (v/v) *n*-hexadecane demonstrated an increased metabolite *m/z* 330.3 formation up to 26.74% (133.7 mg/L). Whereas the OD₆₃₀ value of the cell suspension stained with iodonitrotetrazolium did not exceed 0.57 before DHA addition, the monitored growth of rhodococci at the end of the experiment was on average by 40% higher compared with other experimental variants.

Table 5

¹H- and ¹³C-NMR data of methyl 5 α -hydroxy-abiet-8,11,13-triene-18-oate **8**.

Position	δ_{C} ¹³ C	DEPT	δ_{H} ¹ H	HMBC
1	31.02	CH ₂	2.03 m 2.11 m	H-1/C-2; H-1/C-3; H-1/C-5; H-1/C-9; H-1/C-20
2	17.49	CH ₂	1.84 m	H-2/C-3; H-2/C-4; H-2/C-10
3	31.53	CH ₂	2.41 m 1.58 m	H-3/C-1; H-3/C-2; H-3/C-4; H-3/C-5; H-3/C-18; H-3/C-19
4	50.62	C	—	
5	75.04	C	—	
6	25.31	CH ₂	2.09 m 2.32 m	H-6/C-4; H-6/C-5; H-6/C-7; H-6/C-8; H-6/C-10
7	25.13	CH ₂	2.92 m 3.06 m	H-7/C-5; H-7/C-8; H-7/C-9
8	133.72	C	—	
9	144.34	C	—	
10	42.22	C	—	
11	123.43	CH	7.14 d (8.2)	H-11/C-7; H-11/C-8; H-11/C-10; H-11/C-13
12	123.70	CH	7.01 brdd (8.2, 1.7)	H-12/C-9; H-12/C-14; H-12/C-15
13	145.01	C	—	
14	125.00	CH	6.94 brd (1.7)	H-14/C-9; H-14/C-12; H-14/C-15
15	32.95	CH	2.85 sept (6.9)	H-15/C-16,17; H-15/C-13; H-15/C-14; H-15/C-12
16 /17	23.42	2CH ₃	1.25 d (6.9)	H-16,17/C-13; H-16,17/C-15; H-16/C-16,17; H-17/C-16,17
18	178.34	C	—	
19	19.22	CH ₃	1.46	H-19/C-3; H-19/C-4; H-19/C-5; H-19/C-18
20	28.08	CH ₃	1.33	H-20/C-3; H-20/C-5; H-20/C-9; H-20/C-10
21	51.45	CH ₃ OO	3.73	H-21/C-18
		OH	3.36	OH/C-4; OH/C-5; OH/C-10

Recorded at 400 (¹H-NMR) and 100 (¹³C-NMR) MHz in CDCl₃; δ in ppm, s – singlet, d – doublet, dd – doublet of doublet, sept – septet, m – multiplet, J values (Hz) are indicated in brackets.

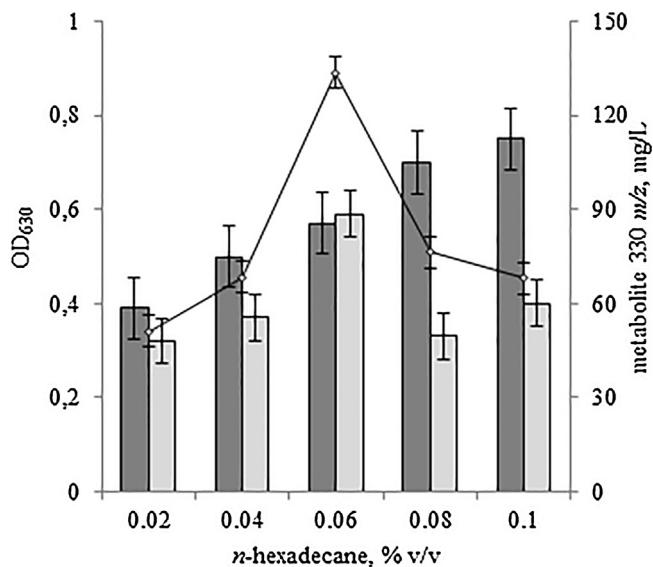


Fig. 7. Concentration (mg/L) of metabolite *m/z* 330.3 (—○—) and growth of *R. erythropolis* IEGM 267 (OD₆₃₀) before DHA addition (■) and after 168 h (□) cultivation with DHA at different initial *n*-hexadecane concentrations (% v/v) in the medium.

3.3. Identification of metabolite *m/z* 330.3

Currently, it is important not only to find effective ways to neutralize toxic DHA, but also to obtain new compounds on its basis. The known DHA derivatives have a wide range of biological activities, i.e. antiulcer [43,44], antimicrobial [45], fungicidal [46,47], anxiolytic [48], antiviral [47,49], and antitumour [44,47,50], and also can be used as intermediates to synthesize bioactive compounds [51]. Numerous examples of biodegradation with the formation of stable metabolites (use of inhibitors) or targeted DHA bioconversions are described in the literature.

For example, 1 β -hydroxy-DHA **2** isolated from the DHA culture medium by fungi *Aspergillus*, *Chephalosporum*, *Cunninghamella*,

Fusarium, *Gibberella* and *Rhizopus* exhibits antimicrobial and inhibitory activities against α -glucosidase [15,16,52]. Selective antitumor and antimicrobial activities are exhibited by 2 α -hydroxy-DHA **3**, the formation of which from DHA is catalyzed by *Mortierella isabellina* and *Mucor rammannianus* [53,54]. 7 β -Hydroxy-DHA **4** formed during DHA biodegradation by *P. abietaniphila* has fungicidal and antitumour activities [13,45,55]. The formation of 15-hydroxy-DHA **5** is catalyzed by *C. elegans*, *G. fujkuroi*, *M. isabellina* and *R. erythropolis*. It has an anti-inflammatory effect and is used as an intermediate to synthesize antiviral and antitumor compounds [15,26,51,53]. 16-Hydroxy-DHA **6** produced in fungal DHA biodegradation inhibits α -glucosidase [15].

The metabolite **8** *m/z* 330.3 (M⁺), is the methyl ether of a novel compound **7** obtained in this study. Methyl 5 α -hydroxy-abiet-8,11,13-triene-18-oate **8** as colorless needle crystal (mp 59.6 °C(*n*-hexane), $[\alpha]_{589}^{22} +24.0$ (c 0.55, CHCl₃), *R*_f value 0.62 (hexane/ethyl acetate 7:3)) was isolated from the methylated extract by *silica gel flash chromatography*. This compound is a hydroxyl-containing DHA derivative, which was confirmed by the presence of the molecular-ion peak at *m/z* 330.3 in the GC-MS and NMR spectral data. So, IR absorption bands at 3480 cm⁻¹ and 1721 cm⁻¹ revealed hydroxyl and carbonyl groups in the structure of DHA derivative **8**. Its ¹³C NMR spectrum (DEPT and APT) as well as ¹³C NMR spectrum of DHA methyl ester [56] indicated the presence of 21 carbon signals, including two tertiary and two secondary methyl groups, a benzene ring, and low-field ester signals (Table 5). Among sp³ carbon signals of compound **8**, DEPT measurements indicated only one tertiary (CH) carbon at δ_{C} 32.95 ppm assigned to C-15 and a new singlet of quaternary carbon at δ_{C} 75.04 ppm assigned to C-5 associated with a hydroxyl group. Alternatively ¹³C NMR spectrum of the initial substrate included two tertiary sp³ carbons C-5 and C-15. The ¹H NMR spectrum of compound **8** showed a new signal for C-5 hydroxyl group at δ_{H} 3.36 ppm and intact septet signal of H-15 at δ_{H} 2.85 ppm, along with the characteristic set of DHA methyl ester signals [57]. Based on 2D ¹H-¹H NMR and ¹³C-¹H NMR spectra (COSY, NOESY, HMQC, and HMBC), compound **8** is suggested to be a methyl abiet-8,11,13-triene-18-oate derivative with a tertiary hydroxyl group at C-5. So, the most informative crosspeaks in

the HMBC spectra were H-1/C-2; H-1/C-3; H-1/C-20; H-3/C-5; H-3/C-18; H-3/C-19; H-6/C-5; H-6/C-7; H-7/C-8; H-7/C-9; H-19/C-4; H-19/C-5; H-19/C-18; H-20/C-10; OH/C-4; OH/C-5; OH/C-10. The NOESY spectrum of compound **8** confirmed the significant NOE interaction between the axial OH group and axial H-3, and weak spatial interactions of OH group with equatorial H-7 and both H-6 atoms, revealing the relative α -orientation of OH at C-5.

4. Conclusion

Based on the screening results of IEGM Collection bioresources, actinobacterial strains capable of complete (*R. rhodochrous* IEGM 107) or partial (*R. erythropolis* IEGM 267) biodegradation of eco-toxic DHA were selected. It was shown that the selected rhodococci were able to degrade DHA at a high initial concentration (500 mg/L) within relatively short period (7 days). Pre-growth of actinobacteria in the presence of *n*-hexadecane is essential for DHA degradation, with optimal *n*-hexadecane concentrations being 0.1 and 0.06% (v/v) for *R. rhodochrous* IEGM 107 and *R. erythropolis* IEGM 267, respectively. The DHA derivative was identified as the new natural product 5 α -hydroxy-abeta-8,11,13-triene-18-oic acid. The data obtained expand our knowledge on catalytic activities of actinobacteria and can be used to develop advanced technologies for effective removal of resin acids from pulp mill effluents.

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References

- [1] S.N. Liss, P.A. Bicho, J.N. Saddler, Microbiology and biodegradation of resin acids in pulp mill effluents: a minireview, *Can. J. Microbiol.* 43 (1997) 599–611, <http://dx.doi.org/10.1139/m97-086>.
- [2] D. Fengel, G. Wegener, *Wood: Chemistry, Ultrastructure, Reactions*, W. de Gruyter, Berlin, 1984.
- [3] S.Z. Chow, D. Shepard, High performance liquid chromatographic determination of resin acids in pulp mill effluent, *Tappi J.* 79 (1996) 173–179.
- [4] G. Peng, J.C. Roberts, Solubility and toxicity of resin acids, *Water Res.* 34 (2000) 2779–2785, [http://dx.doi.org/10.1016/S0043-1354\(99\)00406-6](http://dx.doi.org/10.1016/S0043-1354(99)00406-6).
- [5] Y. Kamaya, N. Tokita, K. Suzuki, Effects of dehydroabietic acid and abietic acid on survival, reproduction, and growth of the crustacean *Daphnia magna*, *Ecotoxicol. Environ. Saf.* 61 (2005) 83–88, <http://dx.doi.org/10.1016/j.ecoenv.2004.07.007>.
- [6] V. Hernández, M. Silva, J. Gavilán, B. Jiménez, R. Barra, J. Becerra, Resin acids in bile samples from fish inhabiting marine waters affected by pulp mill effluents, *J. Chil. Chem. Soc.* 53 (2008) 1718–1721.
- [7] J.K. Volkman, D.G. Holdsworth, D.E. Richardson, Determination of resin acids by gas chromatography and high-performance liquid chromatography in paper mill effluent, river waters and sediments from the upper Derwent Estuary, Tasmania, *J. Chromatogr.* 643 (1993) 209–219, [http://dx.doi.org/10.1016/0021-9673\(93\)80555-M](http://dx.doi.org/10.1016/0021-9673(93)80555-M).
- [8] A. Pérez-de-Mora, E. Madejón, F. Cabrera, F. Buegger, R. Fuß, K. Pritsch, M. Schloter, Long-term impact of acid resin waste deposits on soil quality of forest areas I. Contaminants and abiotic properties, *Sci. Total. Environ.* 406 (2008) 88–98, <http://dx.doi.org/10.1016/j.scitotenv.2008.07.035>.
- [9] M.P. Sobolevskaya, I.A. Lipko, O.P. Moiseenko, V.V. Parfenova, S.S. Afyatullov, Fatty-acid composition of several lake baikal streptomyces, *Chem. Nat. Compd.* 47 (2012) 880–882, <http://dx.doi.org/10.1007/s10600-012-0093-8>.
- [10] M.S. Costa, A. Rego, V. Ramos, T.B. Afonso, S. Freitas, M. Preto, V. Lopes, V. Vasconcelos, C. Magalhães, P.N. Leão, The conifer biomarkers dehydroabietic and abietic acids are widespread in Cyanobacteria, *Sci. Rep.* 6 (2016) 1–11, <http://dx.doi.org/10.1038/srep23436>.
- [11] V.J.J. Martin, Z. Yu, W.W. Mohn, Recent advances in understanding resin acid biodegradation: microbial diversity and metabolism, *Arch. Microbiol.* 172 (1999) 131–138, <http://dx.doi.org/10.1007/s002030050752>.
- [12] C.A. Morgan, R.C. Wyndham, Isolation and characterization of resin acid degrading bacteria found in effluent from a bleached kraft pulp mill, *Can. J. Microbiol.* 42 (1996) 423–430, <http://dx.doi.org/10.1139/m96-058>.
- [13] D.J. Smith, V.J.J. Martin, W.W. Mohn, A cytochrome P450 involved in the metabolism of abietane diterpenoids by *Pseudomonas abietaniphila* BKME-9, *J. Bacteriol.* 186 (2004) 3631–3639, <http://dx.doi.org/10.1128/JB.186.11.3631-3639.2004>.
- [14] P.A. Bicho, V. Martin, J.N. Saddler, Growth, induction, and substrate-specificity of dehydroabietic acid-degrading bacteria isolated from a kraft mill effluent enrichment, *Appl. Environ. Microbiol.* 61 (1995) 3245–3250.
- [15] M.I. Choudhary, M. Atif, S.A. Ali Shah, S. Sultan, S. Erum, S.N. Khan, Atta-Ur-Rahman, biotransformation of dehydroabietic acid with microbial cell cultures and α -glucosidase inhibitory activity of resulting metabolites, *Int. J. Pharm. Pharm. Sci.* 6 (2014) 375–378.
- [16] S.C. Gouiric, G.E. Feresin, A.A. Tapia, P.C. Rossomando, G. Schmeda-Hirschmann, D.A. Bustos, 1 β ,7 β -Dihydroxydehydroabietic acid, a new biotransformation product of dehydroabietic acid by *Aspergillus niger*, *World J. Microbiol. Biotechnol.* 20 (2004) 281–284, <http://dx.doi.org/10.1023/B:WIBI.0000023834.60165.79>.
- [17] T.A. Van Beek, F.W. Claassen, J. Dorado, M. Godejohann, R. Sierra-Alvarez, J.B.P.A. Wijnberg, Fungal biotransformation products of dehydroabietic acid, *J. Nat. Prod.* 70 (2007) 154–159, <http://dx.doi.org/10.1021/np060325e>.
- [18] R. Côté, C. Otis, Étude de la biodégradation de l'acide dehydroabietique par *Bacillus psychrophilus*, *Rev. Des. Sci. L'eau* 2 (1989) 313–324, <http://dx.doi.org/10.7202/705033ar>.
- [19] Z. Yu, W.W. Mohn, Isolation and characterization of thermophilic bacteria capable of degrading dehydroabietic acid, *Can. J. Microbiol.* 45 (1999) 513–519.
- [20] V.J.J. Martin, W.W. Mohn, A novel aromatic-ring-hydroxylating dioxygenase from the diterpenoid-degrading bacterium *Pseudomonas abietaniphila* BKME-9, *J. Bacteriol.* 181 (1999) 2675–2682.
- [21] V.V. Grishko, E.V. Tarasova, I.B. Ivshina, Biotransformation of betulin to betulone by growing and resting cells of the actinobacterium *Rhodococcus rhodochrous* IEGM 66, *Process Biochem.* 48 (2013) 1640–1644, <http://dx.doi.org/10.1016/j.procbio.2013.08.012>.
- [22] T.I. Kylosova, A.A. Elkin, V.V. Grishko, I.B. Ivshina, Biotransformation of prochiral sulfides into (R)-sulfoxides using immobilized *Gordonia terrae* IEGM 136 cells, *J. Mol. Catal. B: Enzym.* 123 (2016) 8–13, <http://dx.doi.org/10.1016/j.molcatb.2015.10.014>.
- [23] M.S. Kuyukina, I.B. Ivshina, *Rhodococcus biosurfactants: biosynthesis, properties and potential applications* Alvarez HM Biol. Rhodococcus. Microbiol. Monogr., 16, Springer, Berlin, 2010, pp. 291–313.
- [24] M.J. Larkin, L.A. Kulakov, C.C.R. Allen, Biodegradation by members of the genus *Rhodococcus*: biochemistry, physiology, and genetic adaptation, *Adv. Appl. Microbiol.* 59 (2006) 1–29, [http://dx.doi.org/10.1016/S0065-2164\(06\)59001-X](http://dx.doi.org/10.1016/S0065-2164(06)59001-X).
- [25] L. Martíneková, B. Uhňáková, M. Pátek, J. Nešvera, V. Křen, Biodegradation potential of the genus *Rhodococcus*, *Environ. Int.* 35 (2009) 162–177, <http://dx.doi.org/10.1016/j.envint.2008.07.018>.
- [26] A.V. Vorob'ev, V.V. Grishko, I.B. Ivshina, E.N. Shmidt, L.M. Pokrovskii, M.S. Kuyukina, G.A. Tolstikov, Microbial transformations of diterpene acids, *Mendeleev Commun.* 11 (2001) 72–73, <http://dx.doi.org/10.1070/MC2001v01n02ABEH001414>.
- [27] K.M. Cheremykh, V.V. Grishko, I.B. Ivshina, Bacterial degradation of ecotoxic dehydroabietic acid, *Catal. Ind.* 9 (2017) 331–338.
- [28] M.S. Kuyukina, I.B. Ivshina, A.Y. Gavrin, E.A. Podorozhko, V.I. Lozinsky, C.E. Jeffree, J.C. Philp, Immobilization of hydrocarbon-oxidizing bacteria in poly(vinyl alcohol) cryogels hydrophobized using a biosurfactant, *J. Microbiol. Methods* 65 (2006) 596–603, <http://dx.doi.org/10.1016/j.mimet.2005.1006>.
- [29] G. Neumann, Y. Veeranagouda, T.B. Karegoudar, Ö. Sahin, I. Mäusezahl, N. Kabelitz, U. Kappelmeyer, H.J. Heipieper, Cells of *Pseudomonas putida* and *Enterobacter* sp. adapt to toxic organic compounds by increasing their size, *Extremophiles* 9 (2005) 163–168, <http://dx.doi.org/10.1007/s00792-005-0431-x>.
- [30] W.W. Mohn, A.E. Wilson, P. Bicho, E.R. Moore, Physiological and phylogenetic diversity of bacteria growing on resin acids, *Syst. Appl. Microbiol.* 22 (1999) 68–78, [http://dx.doi.org/10.1016/S0723-2020\(99\)80029-0](http://dx.doi.org/10.1016/S0723-2020(99)80029-0).
- [31] Z. Yu, W.W. Mohn, Bioaugmentation with resin-acid-degrading bacteria enhances resin acid removal in sequencing batch reactors treating pulp mill effluents, *Water Res.* 35 (2001) 883–890, [http://dx.doi.org/10.1016/S0043-1354\(00\)00353-3](http://dx.doi.org/10.1016/S0043-1354(00)00353-3).
- [32] J.F. Biellmann, G. Branlant, Dégradation bactérienne de l'acide déhydroabiotique par *Flavobacterium resinorum*, *Tetrahedron* 29 (1973) 1227–1236.
- [33] S. Viggør, J. Juhanson, M. Jöesaar, M. Mitt, J. Truu, E. Vedler, A. Heinaru, Dynamic changes in the structure of microbial communities in Baltic Sea coastal seawater microcosms modified by crude oil, shale oil or diesel fuel, *Microbiol. Res.* 168 (2013) 415–427.
- [34] A.M. Warhurst, C.A. Fewson, Biotransformations catalyzed by the genus *Rhodococcus*, *Crit. Rev. Biotechnol.* 14 (1994) 29–73.
- [35] A.A. Elkin, T.I. Kylosova, V.V. Grishko, I.B. Ivshina, Enantioselective oxidation of sulfides to sulfoxides by *Gordonia terrae* IEGM 136 and *Rhodococcus rhodochrous* IEGM 66, *J. Mol. Catal. B: Enzym.* 89 (2013) 82–85.
- [36] I.B. Ivshina, V.V. Grishko, E.M. Nogovitsina, T.P. Kukina, G.A. Tolstikov, Bioconversion of β -sitosterol and its esters by actinobacteria of the genus *Rhodococcus*, *Appl. Biochem. Microbiol.* 41 (2005) 551–557.

- [37] W.W. Mohn, Z. Yu, E.R.B. Moore, A.F. Muttray, Lessons learned from *Sphingomonas* species that degrade abietane triterpenoids, *J. Ind. Microbiol. Biotechnol.* 23 (1999) 374–379, <http://dx.doi.org/10.1038/sj/jim/2900731>.
- [38] M.S. Kuyukina, I.B. Ivshina, I.O. Korshunova, E.V. Rubtsova, Assessment of bacterial resistance to organic solvents using a combined confocal laser scanning and atomic force microscopy (CLSM/AFM), *J. Microbiol. Methods* 107 (2014) 23–29, <http://dx.doi.org/10.1016/j.mimet.2014.08.020>.
- [39] I.B. Ivshina, M.S. Kuyukina, J.C. Philp, N. Christofi, Oil desorption from mineral and organic materials using biosurfactant complexes produced by *Rhodococcus* species, *World J. Microbiol. Biotechnol.* 14 (1998) 711–717, <http://dx.doi.org/10.1023/A:1008885309221>.
- [40] E.V. Tarasova, V.V. Grishko, I.B. Ivshina, Cell adaptations of *Rhodococcus rhodochrous* IEGM 66 to betulin biotransformation, *Process Biochem.* 52 (2017) 1–9, <http://dx.doi.org/10.1016/j.procbio.2016.10.003>.
- [41] P. Atrat, P. Hosel, W. Richter, H.W. Meyer, C. Horhold, Interactions of *Mycobacterium fortuitum* with solid sterol substrate particles, *J. Basic Microbiol.* 31 (1991) 413–422, <http://dx.doi.org/10.1002/jobm.3620310605>.
- [42] M.S. Kuyukina, I.B. Ivshina, I.O. Korshunova, G.I. Stukova, A.V. Krivoruchko, Diverse effects of a biosurfactant from *Rhodococcus ruber* IEGM 231 on the adhesion of resting and growing bacteria to polystyrene, *AMB Express* 6 (2016) 14, <http://dx.doi.org/10.1186/s13568-016-0186-z>.
- [43] X.H. Lin, M.N. Cao, W.N. He, S.W. Yu, D.A. Guo, M. Ye, Biotransformation of 20(R)-panaxadiol by the fungus *Rhizopus chinensis*, *Phytochemistry* 105 (2014) 129–134, <http://dx.doi.org/10.1016/j.phytochem.2014.06.001>.
- [44] H. Wada, S. Kidato, M. Kawamori, T. Morikawa, H. Nakai, M. Takeda, S. Saito, Y. Onoda, H. Tamaki, Antiulcer activity of dehydroabietic acid derivatives, *Chem. Pharm. Bull.* 33 (1985) 1472–1487, <http://dx.doi.org/10.1248/cpb.37.3229>.
- [45] S. Savluchinske-Feio, M.J.M. Curto, B. Gigante, J.C. Roseiro, Antimicrobial activity of resin acid derivatives, *Appl. Microbiol. Biotechnol.* 72 (2006) 430–436, <http://dx.doi.org/10.1007/s00253-006-0517-0>.
- [46] S. Savluchinske-Feio, L. Nunes, P.T. Pereira, A.M. Silva, J.C. Roseiro, B. Gigante, M.J. Marcelo Curto, Activity of dehydroabietic acid derivatives against wood contaminant fungi, *J. Microbiol. Methods* 70 (2007) 465–470, <http://dx.doi.org/10.1016/j.mimet.2007.06.001>.
- [47] M.A. González, D. Pérez-Guaita, J. Correa-Royer, B. Zapata, L. Agudelo, A. Mesa-Arango, L. Betancur-Galvis, Synthesis and biological evaluation of dehydroabietic acid derivatives, *Eur. J. Med. Chem.* 45 (2010) 811–816, <http://dx.doi.org/10.1016/j.ejmech.2009.10.010>.
- [48] I.A. Tolmacheva, A.V. Tarantin, A.A. Boteva, L.V. Anikina, Y.B. Vikharev, V.V. Grishko, A.G. Tolstikov, Synthesis and biological activity of nitrogen-containing derivatives of methyl dehydroabietate, *Pharm. Chem. J.* 40 (2006) 489–493, <http://dx.doi.org/10.1007/s11094-006-0161-0>.
- [49] J.R. Tagat, D.V. Nazarene, M.S. Puar, S.W. Mccombe, A.K. Ganguly, Synthesis and anti-herpes activity of some a-ring functionalized dehydroabietane derivatives, *Bioorg. Med. Chem. Lett.* 4 (1994) 1101–1104.
- [50] X.C. Huang, M. Wang, Y.M. Pan, G.Y. Yao, H.S. Wang, X.Y. Tian, J.K. Qin, Y. Zhang, Synthesis and antitumor activities of novel thiourea α -aminophosphonates from dehydroabietic acid, *Eur. J. Med. Chem.* 69 (2013) 508–520, <http://dx.doi.org/10.1016/j.ejmec.2013.08.055>.
- [51] M. González, Aromatic abietane diterpenoids: their biological activity and synthesis, *Nat. Prod. Rep.* (2015) 684–704, <http://dx.doi.org/10.1039/C4NP00110A>.
- [52] A.A. Tapia, M.D. Vallejo, S.C. Gouiric, G.E. Feresin, D.A. Rossomando, Pedro C. Bustos, Hydroxylation of dehydroabietic acid by *Fusarium* species, *Phytochemistry* 46 (1997) 131–133.
- [53] J.P. Kutney, L.S.L. Choi, G.M. Hewitt, P.J. Salisbury, M. Singh, Biotransformation of dehydroabietic acid with resting cell suspensions and calcium alginate-immobilized cells of *Mortierella isabellina*, *Appl. Environ. Microbiol.* 49 (1985) 96–100.
- [54] O. Ozsen, I. Kiran, I. Dag, O. Atli, G.A. Cifci, F. Demirci, Biotransformation of abietic acid by fungi and biological evaluation of its metabolites, *Process Biochem.* 52 (2017) 130–140, <http://dx.doi.org/10.1016/j.procbio.2016.09.022>.
- [55] N.Y. Yang, L. Liu, W.W. Tao, J.A. Duan, L.J. Tian, Diterpenoids from *Pinus massoniana* resin and their cytotoxicity against A431 and A549 cells, *Phytochemistry* 71 (2010) 1528–1533, <http://dx.doi.org/10.1016/j.phytochem.2010.06.008>.
- [56] M.S. de Carvalho, L.H.B. Baptista, P.M. Imamura, 13C and 1H NMR signal assignments of some new synthetic dehydroabietic acid derivatives, *Magn. Reson. Chem.* 46 (2008) 381–386, <http://dx.doi.org/10.1002/mrc.2187>.
- [57] L.L. Landucci, D.F. Zinkel, Prediction of the 1H and 13C NMR spectra of the abietadienoic acid, *Holzforschung* 45 (1991) 341–346.